1

# **Texture Properties and Kinetic Parameters Associated to Carbon Materials Obtained from Sawdust of Algarroba Wood. 1. Application in Phenol Photodetoxification**

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**Abstract:** Carbon materials were obtained by pyrolysis of the sawdust of Algarroba wood under  $CO_2$  or  $N_2$  atmospheres at temperatures from 200°C up to 900°C. Carbon materials were characterized by surface and micropore areas, volume pore and mean pore diameter. An approach of the kinetic parameters (apparent constant-rates and activation energies) associated with the synthesis of these carbon materials is also presented. From the kinetic data of thermal degradation of Algarroba wood, three different pyrolysis zones were verified. First, an incipient carbonization between 200 up to 350°C, second an extensive carbonization between 350 up to 600°C, and a third step involving the pore formation at temperatures higher than 600°C. As a target application of carbon materials, results of phenol adsorption and phenol photodegradation under UV-irradiation indicated the potential of carbon materials in presence of TiO<sub>2</sub> for treatment of polluted waters.

Keywords: Thermal degradation, Carbon materials, Kinetics, Activation energy, Photocatalysis.

# **1. INTRODUCTION**

Carbon-based materials as activated carbon (AC) have been extensively used as adsorbents and catalytic supports mainly due to the high surface area up to  $3000m^2 g^{-1}$  and a wide range of pore sizes [1]. Bituminous and lignocelluloses are the most common precursors employed as starting materials of activated carbons. Adsorption on AC depend on two fundamental characteristics, accessible internal surface area where adsorption takes place, and the presence of sites at which chemisorptions may occur. It is important to have a complete knowledge of the activated carbon properties because besides conventional applications, there is an increasing interest in develop AC for specifics needs. For example, AC has been employed in catalytic heterogeneous reactions such as hydrogasification [2, 3], hydrodesulphurization [4, 5] and photocatalytic detoxification of waste waters [6-8], where its application is directly associated with textural and surface properties of carbon [9, 10]. Also, an increasing interest in the influence of wood char [11] and wood-derived carbon-supported catalysts on the kinetics of methane decomposition [12] and dry methane reforming [13, 14] have been performed. Consequently, the study the kinetics parameters related with the synthesis of carbons from lignocelluloses precursor materials, such as sawdust from wood have receive some attention [15, 16]. In addition, rigorous kinetic studies about the thermal decomposition of biomass materials [17] and specifically about cellulose pyrolysis [18] have permitted to clarify the influence of gas evolution and the mechanisms related with the pyrolysis of lignocellulosic materials. In this sense, the main objective of this work is focussed to the synthesis and textual characterization of carbons materials obtained from the pyrolysis of sawdust of Algarroba wood under CO<sub>2</sub> and N<sub>2</sub> atmospheres. An additional study of the kinetics parameters and energies associated with the synthesis of these carbon materials is also included as an approach to insight into the possible mechanistic steps of the wood activation and to correlate these kinetics parameters with the textural properties. Finally, the influence of two carbons upon the phenol adsorption and on the photoefficiency of  $TiO_2$  in the phenol photo detoxification under UV-irradiation was studied as a targeted application.

# 2. EXPERIMENTAL

# 2.1. Raw Material and Activated Carbon Characterization

Sawdust from Algarroba (*Hymenaea Courbaril*) wood was employed. This wood was selected because in addition to low ash content it is a hard wood and therefore useful to some applications. For example, Laine *et al.* [19] has employed Algarroba wood to obtain activated carbons filters in the shape of monoliths with a high potential application in the treatment of polluted water as we showed in an early work [20]. Algarroba's sawdust was firstly crashed and sieved [21] before characterization and pyrolysis studies. The mean size of particles was 350µm; however, an important proportion of particles with size lower than this was noticed that means that AC prepared will be on the micro particle solid classification. Tests of moisture at 120°C by 2h

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Fig. (1). Scheme of activation set-up.

[22], ash at 550°C by 14h [23] and volatiles at 850°C by 7min [24] were also performed following ASTM standards [22-24]. Bulk density was also estimated by geometrical filled method with sieved microparticles of 350µm. Identification and quantification of the elements Al, B, Ba, Cr, Cu, Mg, Mn, Na, Ni, P, Si, Sr, Zn was performed in an inductively-coupled plasma atomic emission spectrophotometer (ICP-AES) from Varian (model 730ES-Axial, series EL07114070). Quantification of these elements was performed by using a multielemental standard from High Purity (E.P.A. 2007) at five different concentrations. The determination of Ca, Fe and K was performed in an atomic absorption (AA) spectrophotometer from Perkin Elmer (model Analyst 400, series 201S3090705). In these cases, quantification was performed by using calibration curves with five different concentrations from standards FisherScientific (for Fe) and Ricca Chemical Company (for Ca and K). Other elements such as As, Be, Cs, Cd, Co, Ge, In, La, Mo, Pb, Sb, Sc, Se, Sn, Ti, Tl, and V, were also explored but they were not detected. Analysis was done by triplicate for the samples and ashes and the reproducibility of results being better than 99%. Textural characterization was performed by N<sub>2</sub> adsorption-desorption isotherms at 77K. The full isotherms in the range of  $4 \times 10^{-3}$  to 84 kPa were measured in a Micromeritics ASAP-2010. Equivalent surface area, micropore area and volume, and pore diameters were obtained by Brunauer-Emmet-Teller (BET), Harkins-Jura (HJ) and Horvath-Kawazoe (HK) methods, respectively [25-27]. HJ and HK were employed because they provide an accounting for the

adsorbed layer on pore walls when calculating pore size distribution and because they are very useful when different types of pores are involved. Therefore, HJ and HK methods are proper for the present case because carbon materials can contain slits and spherical pores.

#### 2.2. Thermal Degradation of Sawdust

Pyrolysis was carried out at nearly constant atmospheric pressure (100kPa) inside a tubular kiln (HM 16, Heraeus GmBH Germany). Fig. (1) shows a schematic representation of the equipment used in the present work where reactor employed was made of stainless steel or Pyrex glass depending of final pyrolysis temperature. Final temperatures of oven were calibrated with standard deviation less than 2% by using a rheostat (Heraeus). Isothermal conditions were revised and oven remains its final temperature during all kinetic studies. Samples of 3g of sieved sawdust were placed in the oven at ambient temperature. Samples were pyrolyzed under CO2 or N2 gas flows (100mL.min<sup>-1</sup>) from 20 up to the final temperatures between 200°C up to 900°C at constant heating rate (10°C/min). Temperature was kept constant for different holding times (0, 1, 2h) after achieve the final selected temperature. From a pre-industrial scale point of view, the controlled pyrolysis of lignocellulosic precursors as sawdust from wood is the most single and cheaper methodology to obtain carbon material. The selection of the temperatures ranges studied in the present work was made to compare against previous results from Laine et al. [4,5,19] and from our own group [9,10]. Different wood samples were used for



**Fig. (2).** Mass remaining (%m) after pyrolysis of Algarroba wood. (A): Under CO<sub>2</sub> flow. (B): Under N<sub>2</sub> flow.

each experiment and changes between initial  $(m_o)$  and mass remaining  $(m_t)$  after these times were verified. From the mass remaining data of pyrolysis as function of the gas atmosphere, the temperature and the time of reaction, the apparent constant rates and activation energies in three different pyrolysis zones were estimated as follows.

## 2.3. Estimation of Kinetic Parameters

The kinetic of thermal decomposition of the wood was studied assuming a first-order reaction mechanism as suggested the mass remaining trends (Fig. 2) obtained after the thermal degradation of the wood. This will be discussed in the result and discussion section. In this sense, first-order apparent constants ( $k_{ap,Ti}$ ) were estimated as a function of temperature from the linear regression of equation (1), giving equation (2):

$$d(\%m) = -k_{ap,Ti} \%m \cdot dt$$
(1)

$$Ln(\%m_o/\%m_t) = k_{ap,Ti} \cdot t$$
<sup>(2)</sup>

where  $\%m_o$  is the mass remaining at zero reaction time,  $\%m_t$  is the mass remaining after 1 and 2h reaction, t is the reaction time and  $k_{ap,Ti}$  is the first-order apparent rate constant at the specific temperature  $T_i$ . Activation energies associated to different pyrolysis zones in the synthesis of carbon materials were estimated from the Arrhenius kinetic formulation [17]:

$$k_{ap,Ti} = A \cdot exp(-Ea/R \cdot T)$$
(3)

$$Ln(k_{ap,Ti}) = Ln(A) - (E_a/R \cdot T)$$
(4)

where  $E_a$  and A in equations 3 and 4 correspond to the activation energy and to the frequency factor, respectively and R corresponds to the gas constant. A is also known as the preexponential Arrhenius factor and is commonly related with the probability of reaction [28]. In the present study, we found that in the pyrolysis zone between 300-600°C, the higher the temperature the lower the first-order apparent rate constant. Therefore, we used a first-order differential equation (Eq. 5), as an approximation to estimate the activation energy by means of the linear transformation of equation 6.

$$dk_{ap,Ti} = (-k_{ap,Ti} \cdot R \cdot dT)$$
(5)

$$Ln(k_{ap,To}/k_{ap,Ti}) = (1/E_a)R \cdot Ti$$
(6)

## 2.4. Phenol Photodetoxification

Phenol was purchased from Aldrich. The photocatalyst was TiO<sub>2</sub> Degussa P25, mainly anatase (ca. 70%) under the shape of non-porous polyhedral particles of ca. 30 nm mean size with a surface area of 50  $m^2/g$ . Activated carbons prepared under CO2 and N2 flow at 800°C and 1h were selected for the study because they showed the highest surface areas (discussed below). These samples were denoted:  $TiO_2$ -AC<sub>CO2</sub> and  $TiO_2$ -AC<sub>N2</sub>. The experimental set-up has been described before [6, 7] but it can be summarized as follows. The batch photoreactor was a cylindrical flask made of Pyrex of ca. 100 ml with a bottom optical window of ca. 4 cm diameter and was open to air. Irradiation was provided by a high pressure mercury lamp (Phillips HPK 125 W) and was filtered by a circulating-water cell (thickness 2.2 cm) equipped with a 340 nm cut-off filter (Corning 0.52). The water cell was used to remove all the IR beams, thus preventing any heating of the suspension, especially in the presence of black activated carbon. The cut-off filter, although decreasing the overall UV-light power available, enables one to eliminate any photochemical side reaction. Millipore disks (0.45 µm) were used to remove particulate matter before HPLC analysis. The HPLC system comprised a LDC/Milton Roy Constametric 3200 isocratic pump and a Waters 486 tunable absorbance detector (Millipore) adjusted at 270 nm for the detection of phenol. The quantity of 50 mg of titania was chosen since in our conditions there is a full absorption of the UV light entering the photoreactor and because it has shown a optimum of composition in the photodegradation of phenol [6]. The quantity of 10 mg AC was chosen to ensure a good adsorption of phenol related to the high surface area of AC without disturbing the UV absorption by titania nor phenol adsorption on it [6]. Samples of the suspension were removed at regular intervals for analysis and from the linear regression of the kinetic data of phenol disappearance as a function of reaction time  $Ln(C_o/C_t) = f(t)$ , the first-order apparent rateconstans (k<sub>app-phenol</sub>) were obtained to compare the photoefficiency of the TiO<sub>2</sub>-AC against TiO<sub>2</sub> alone. Further details can be verified elsewhere [6].

#### **3. RESULTS AND DISCUSSION**

#### 3.1. Raw Material Characterization

Results of wood characterization are shown in Table 1. It can be inferred that moisture (about 10% wt water content) would not affect activation processes. Volatile high content (about 72% wt) indicates that Algarroba's samples of the

Table 1. Characterization of Hymenaea Courbaril Wood

Moisture (wt%)	$9.8 \pm 0.4$
Ash (wt%)	$1.30\pm0.04$
Volatiles (wt%)	$72.1\ \pm 0.7$
Fixed carbon (wt%)	$16.8 \pm 1,3$
Apparent Density (g/cm <sup>3</sup> )	$0.89\pm0.03$

present work are from a young wood and therefore, low yields should be obtained after pyrolysis. A low metal content (about 1.3%wt) was also detected in the carbonaceous precursor. Table 2 shows a summary of the atomic absorption (AA) analysis obtained from ashes of Algarroba. In general, low metal content of about 1.3wt% were detected in the composition of the wood suggesting that carbon materials would have a high purity. These metals were mainly alkaline (Na and K), and alkaline-earth (Ca, Mg, Ba and Sr). Also, it must be note an important quantity of phosphor (ca. 0.7% wt). Laine et al. [19] have reported that P found in carbons obtained from Algarroba is commonly structured in the shape of cyclic or linear polyphosphates  $[(P_nO_{3n})]n$ - and  $[(P_nO_{3n+1})](n + 2)$ -, respectively, that can be coordinated to the carbon sheets indicating a strong interaction between phosphates anions and carbon atoms. Low content of amphoteric elements (Al and Si) and first row transition metals (Cr, Cu, Fe, Ni, Zn and Mn) were also detected.

## 3.2. Thermal Degradation of Wood

In preliminary studies (not shown) from 300°C up to 450°C under CO<sub>2</sub> flow, besides 10°C/min we employed 5°C/min but no representative changes were detected on mass remaining. For that reason we selected 10°C/min in order to compare the present yields against those reported previously by Laine et al. [19] for monolithic carbons at low temperature and by our group for activated carbons prepared at high temperatures [20]. The mass remaining (%m) after thermal degradation under CO<sub>2</sub> or N<sub>2</sub> flow are show in Fig. (2A and 2B), respectively. This figure also includes the moisture lost by the heating up to 120°C. In both cases, it can be noted that thermal decomposition started about 200°C in agreeing with previous works of lignocelluloses materials decomposition [15, 16, 29]. Fig. (2) shows three different thermal degradation zones as a function of temperature. The first zone can be distinguish between 200-300°C, the second one between 350-600°C and the last one between 600-900°C. For the case of thermal degradation under CO<sub>2</sub> atmosphere, it can be note from Fig. (2A) that the higher the activation temperature and reaction time the lower the mass remaining. For the case of pyrolysis under  $N_2$  flow, Fig. (2B) shows a similar behaviour up to 700°C where the mass remaining (%m) is practically constant at higher temperatures than 700°C. The different trends in the mass remaining as a function of gaseous atmospheres was expected because carbonized wood reacts efficiently with CO<sub>2</sub> at temperature higher than 690°C by the well-known Boudouart reaction: [30]

	Table 2. A	Atomic Absor	ption Analysis	of Ashes from	Algarroba Wood
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Element	Concentration in Ashes (mg/L)	Proportion (%)	Wood Composition (wt%)
Al	18.31	1.05	0.0136
Ba	0.51	0.03	0.0004
В	1.83	0.10	0.0014
Ca	538.46	30.80	0.4066
Cr	0.17	0.01	0.0001
Cu	1.58	0.09	0.0012
Fe	86.46	4.95	0.0653
Mg	112.05	6.41	0.0846
Mn	0.46	0.03	0.0003
Ni	0.15	0.01	0.0001
Р	934.57	53.47	0.7057
К	42.58	2.44	0.0322
Si	0.82	0.05	0.0006
Na	2.61	0.15	0.0020
Sr	2.69	0.15	0.0020
Zn	4.66	0.27	0.0035
Total	1747.91	100.00	1.30

	Zero Time	Residence	1h Res	idence
Temp. (°C)	$S_{BET} CO_2 (m^2/g)$	$S_{BET} N_2 \left(m^2 / g\right)$	$S_{BET} \ CO_2 \ (m^2/g)$	$S_{BET}  N_2  (m^2\!/g)$
350	45 ± 3	7 ± 1	92 ± 1	$34 \pm 1$
450	171 ± 3	23 ± 1	$350\pm7$	$220\pm 6$
600	$288\pm 6$	$310\pm 8$	870 ± 17	497 ± 1
700	$330\pm7$	$328\pm 6$	$1038\pm28$	$527 \pm 2$
800	426 ±19	$372 \pm 7$	1167 ± 31	$549 \pm 2$
900	534 ± 10	$387 \pm 8$	$752 \pm 20$	471 ± 12

#### Table 3. B.E.T Surface Area of Carbon Materials Prepared by Pyrolysis Under CO<sub>2</sub> or N<sub>2</sub> Flow

Table 4. Micropore Area (µpore<sub>area</sub>), Micropore Volume (µpore<sub>volume</sub>) and Mean width of Pore (W<sub>pore</sub>) of Carbon Materials Prepared by Pyrolysis Under CO<sub>2</sub> or N<sub>2</sub> Flow after 1h Reaction

	CO <sub>2</sub> Flow	CO <sub>2</sub> Flow	CO <sub>2</sub> Flow	N <sub>2</sub> Flow	N <sub>2</sub> Flow	N <sub>2</sub> Flow
Τ (°C)	µpore <sub>area</sub> (m²/g) <sup>a</sup>	μpore <sub>volume</sub> (cm³/g) <sup>b</sup>	W <sub>pore</sub> (Å) <sup>c</sup>	µpore <sub>area</sub> (m²/g) <sup>a</sup>	µpore <sub>volume</sub> (cm³/g) <sup>b</sup>	W <sub>pore</sub> (Å) <sup>c</sup>
350°C	<sup>d</sup>	<sup>d</sup>	1815	<sup>d</sup>	<sup>d</sup>	922
450°C	327	0.010	54.5	202	0.101	43.3
600°C	624	0.241	20.6	449	0.172	16.3
700°C	892	0.411	20.0	494	0.188	15.8
800°C	1003	0.462	18.6	515	0.195	15.8
900°C	646	0.573	6.72	455	0.223	5.33

<sup>a,b</sup>Obtained by HJ method. <sup>c</sup>Obtained by HK method. <sup>d</sup>Not detected

$$C + CO_2 \rightarrow 2CO$$
 (7)

which is a bimolecular process. By contrast, under  $N_2$  atmosphere the unimolecular thermal decomposition of wood occurs from low temperatures as 200°C up to 700°C. At higher temperatures than 700°C and under inert atmospheres, the aromatization of carbon atoms has been previously reported by us and other authors [31-35]. This proposal is in agreement with the constancy of the mass remaining at temperatures higher than 700°C (Fig. **2B**).

## 3.3. Texture of AC

A summary of BET surface areas ( $S_{BET}$ ), micropore area ( $\mu$ pore<sub>area</sub>), micropore volume ( $\mu$ pore<sub>volume</sub>) and mean pore width ( $W_{pore}$ ) corresponding to carbon materials prepared under CO<sub>2</sub> and N<sub>2</sub> atmospheres are compiled in Table **3** and **4**. For the samples heated up to final temperature and with zero reaction time, it can be seen from Table **3** that the higher the temperatures of thermal degradation of wood the higher the S<sub>BET</sub>. This can be associated to a pore opening by the thermal effect [31, 35]. By contrast, for the case of 1h reaction time, maxima values in BET surface area were obtained at 800°C for both atmospheres. This trend has been previously reported by some of us in the activation of sac-

charose under CO2 and N2 flow [31] and can be observed in Fig. (3). In the same direction as above, the higher the time of thermal degradation of wood the higher the S<sub>BET</sub>. This effect is consequence of a pore opening by the evolution of gas species for the case of the pyrolysis under N2 flow or by an enhance interaction between the carbon layers with CO<sub>2</sub> induced by the thermal effect [31, 35]. In consequence, BET surface area of carbon materials prepared by 1h of thermal treatment were higher than those prepared only heating up to the final temperature. For this reason, carbons prepared by 1h of thermal degradation were selected for the porosimetry study and results are showed in Table 4. In general, the higher the temperature of degradation of the wood the higher the micropores volume (µporevolume) of carbons prepared. However, for the micropore area (µporearea) parameter, a maxima value is reached at 800°C in agreeing with SBET (Table 3) and in most of cases, the microporous area contributes with about 90% of the total surface area. It can be seen in Table 4 that the higher the final temperature of thermal degradation the lower the mean width of pore (W<sub>pore</sub>). For temperatures of 350 and 450°C macro- and mesoporous carbons were obtained respectively, whereas between 600 and 800°C microporous materials were obtained. Finally, ultramicroporous carbon materials were obtained at 900°C with the lowest mean width of pore of about 6.7Å and 5.3Å,





Fig. (3). Trend of BET surface area as a function of temperature for 1h activation under CO<sub>2</sub> flow or 1h of pyrolysis under N<sub>2</sub> flow.

Table 5. Linear Regression of Kinetic Data from Fig. (2).

Temperature (°C)	CO <sub>2</sub> Flow k <sub>ap</sub> (min <sup>-1</sup> )	$\mathbf{F}^{\mathbf{a}}$	N <sub>2</sub> Flow k <sub>ap</sub> (min <sup>-1</sup> )	$\mathbf{F}^{a}$
200	2.294x10 <sup>-4</sup>	0.99	1.420x10 <sup>-4</sup>	0.80
250	6.031x10 <sup>-4</sup>	0.85	8.372x10 <sup>-4</sup>	0.87
300	23.38x10 <sup>-4</sup>	0.99	27.73x10 <sup>-4</sup>	0.99
350	9.564x10 <sup>-4</sup>	0.94	13.38x10 <sup>-4</sup>	0.99
450	6.714x10 <sup>-4</sup>	0.90	7.609x10 <sup>-4</sup>	0.90
600	3.797x10 <sup>-4</sup>	0.87	4.563x10 <sup>-4</sup>	0.98
700	9.835x10 <sup>-4</sup>	0.90	3.429x10 <sup>-4</sup>	0.81
800	36.95x10 <sup>-4</sup>	0.99	4.073x10 <sup>-4</sup>	1.00
900	252.4x10 <sup>-4</sup>	0.96	7.597x10 <sup>-4</sup>	1.00

<sup>a</sup>F corresponds to the square factor of linear regression

under  $CO_2$  and  $N_2$  atmospheres, respectively. It is important to remark that in all temperatures, the mean width of pores were lower for the carbons obtained under  $N_2$  flow than those obtained under  $CO_2$  atmosphere. For example, at 350°C, the mean width of carbon obtained under  $N_2$  flow was about the half than that obtained under  $CO_2$  flow, 922Å against 1815Å, respectively. This trend was the same at 450°C where mesoporous carbon materials were obtained, 43Å against 54Å, under  $N_2$  and  $CO_2$  atmospheres, respectively, and also the same in the range (600-900°C) where carbons obtained were microporous. In short, these results suggest that pore opening is influence by reaction atmosphere and this fact could be related with the activation energies associated with the synthesis of carbon materials discussed as follows.

## 3.4. Kinetic Parameters

We have pointed out in the experimental section that the kinetic of thermal decomposition of sawdust of Algarroba wood was studied assuming a first-order reaction mechanism. In this sense, first-order apparent constants ( $k_{ap,Ti}$ ) were estimated as a function of temperature by equations 1 and 2.  $k_{ap,Ti}$  values in Table **5** and mass remaining in Fig. (**2**), sug-

	CO <sub>2</sub> flow		CO <sub>2</sub> flow		N <sub>2</sub> Flow	w
Temperature (°C)	A (min <sup>-1</sup> )	E <sub>a</sub> (kcal.mol <sup>-1</sup> )	A (min <sup>-1</sup> )	E <sub>a</sub> (kcal.mol <sup>-1</sup> )		
200-300	112.5	12.4	3908.1	16.1		
350-600	23.54	0.09	33.73	0.06		
600-900	2460	27.8	0.03	8.84		

Table 6. Arrhenius Parameters as Function of Gas Atmosphere and Pyrolysis Zones

gest that mechanism involved in the synthesis of the present carbon materials would be composed by 3 different pyrolysis zones. Firstly, an incipient carbonization process take place at low temperature (200-300°C) by consequence of carbohydrate dehydration (CH<sub>2</sub>O  $\rightarrow$  C + H<sub>2</sub>O) to produce macroporous materials. Second, in a middle pyrolysis zone at temperatures between 350 up to 600°C, an extensive carbonization of sawdust take place to produce mesoporous materials. The third pyrolysis zone is located at high temperatures (T > T)600°C) where the microporous formation take place by a rearrangement of carbon layers under N2 atmosphere or by gasification according to the Boudouart reaction discussed above [30]. This step would be influenced by some of the wood composition elements such as K, Mg and Ca, because they could act as catalysts in the gasification reaction [33, 34]. It can be seen from Table 5 that at the low pyrolysis zone (200-300°C) and at the middle pyrolysis zone (350-600°), k<sub>ap,Ti</sub> values obtained under N<sub>2</sub> flow are slightly higher than those obtained under CO<sub>2</sub> flow. This could be related with the fact that at low and middle temperatures, recondensation phenomena of volatile matter on solid would inhibit the evolution of tars in presence of CO<sub>2</sub> [30]. Also, according to the literature [35], CO<sub>2</sub> treatment may produce oxygen atomic bridges to form slit micropores between graphene sheets. By contrast, at the pyrolysis zone of highest temperatures, k<sub>ap.Ti</sub> are clearly higher under CO<sub>2</sub> atmosphere than those obtained under N2 flow. This could be due to the most efficient interaction between CO<sub>2</sub> and carbon because at temperatures higher than 700°C Boudouart reaction is at thermodynamic regime's condition [30]. Energies associated to the pyrolysis zones of lower and higher temperatures, 200-350°C and 600-900°C, respectively, were estimated by the equations 3-4 and these values are showed in Table 6. It must be note that in the middle pyrolysis zone (350-600°),  $k_{ap,Ti}$  values in Table 5 indicate that the higher the temperature the lower the k<sub>ap,Ti</sub>. As we indicate above in the experimental section, the conventional Arrhenius kinetic formulation cannot be employed in this middle zone and therefore equation 5 and 6 were employed to estimate the activation energies. We discuss firstly results obtained at the low and high ranges of temperatures. It can be seen from values in Table 6 that in the pyrolysis zone of low temperatures (200-300°C), activation energy obtained under CO<sub>2</sub> flow is smaller than that obtained under N<sub>2</sub> flow (12.4kcal.mol<sup>-1</sup> against 16.1kcal.mol<sup>-1</sup>). This can be due to the fact that  $CO_2$ is an oxidizing gas that reacts with carbonaceous precursor inducing the evolution of volatiles while under inert gas as N<sub>2</sub>, a higher energy quantity is required to induce volatile evolution reactions. By contrast, in the pyrolysis zone between 600°C and 900°C, the activation energy obtained under CO<sub>2</sub> flow is about three times higher than that obtained under  $N_2$  flow (27.8kcal.mol<sup>-1</sup> against 8.8kcal.mol<sup>-1</sup>). This result can be explained by the fact that under N<sub>2</sub> atmosphere the mass remaining at temperature higher than 700°C is practically constant (Fig. 2B). Therefore, it seems to be logical to suggest that energy requirement are lower for the activation under N<sub>2</sub> flow by means of carbon atoms re-arrangement that can induce the formation under inert atmosphere of slit microporous as discussed above. By contrast, the energy requirements for the carbon gasification under CO<sub>2</sub> flow are higher [32]. This result was expected because at 900°C, the wood was fully carbonized without any evident changes in the mass remaining and therefore, the interaction between this material enrichment in carbon atoms and N<sub>2</sub> is negligible. This discussion is reinforced by the results of micropore volume from Table 4, which show that at temperatures higher than 600°C, the production of micropore volume in the structure of the carbons is remarkably higher under CO<sub>2</sub> flow than that obtained under N2 flow. Finally, it is important to remark that in the pyrolysis zone of middle temperatures (350-600°C), values of activation energies (Table 6) obtained on CO2 or N2 atmospheres were very similar and both practically negligible (0.09 and 0.06kcal.mol<sup>-1</sup>) which reinforce the above discussion which suggest that in this pyrolysis zone, an incipient aromatization of carbon atoms can take place independently of reaction atmosphere [1]. Regarding to frequency pre-exponential factor (A), it can be seen from values in Table 6 that in the pyrolysis zones of low and high temperatures its trends are very similar than those followed by the activation energies. Frequency factor values obtained at the pyrolysis zone (350-600°C), and mainly those obtained under N<sub>2</sub> atmosphere at the pyrolysis zone (600-900°C) are clearly lower than those commonly obtained in uni- or bimolecular homogeneous gas phase reactions [28] indicating that pyrolysis of sawdust of Algarroba wood occurs in heterogeneous phase. The very low value of A factor obtained under  $N_2$  atmosphere (3.1x10<sup>-2</sup>min<sup>-1</sup>) in the pyrolysis zone of higher temperatures is in agreement with the above suggestion regarding to the re-arrangement of carbons layers without no representatives lost of mass (Fig. 2B). Present efforts are aimed to verify this behaviour with other kinds of carbon precursor materials and to seek evidences to confirm the three steps proposed for the pyrolysis of wood as and the influence of composition upon kinetic parameters.

### 3.5. General Discussion

According to Bahng and co-workers [36] the pyrolysis of biomass can be described, assuming a single reaction as following Eq. (7) [37-40]:

$$d\alpha/dt = A \cdot f(\alpha) \cdot \exp(-E/R \cdot T)$$
(7)

The variable  $\alpha$  is the degree of transformation,  $\alpha = (m_0 - m)/(m_0 - m_{\infty})$ , where m is the mass of biomass and the subscripts 0 and  $\infty$  refer to the initial and residual amounts, respectively. Parameters E and A are the activation energy and the pre-exponential factor and R is the ideal gas constant and  $[f(\alpha) = 1 - \alpha]$ , assuming a one-step first-order reaction (shown in Eq. 8).

$$d\alpha/dt = A \cdot (1-\alpha) \cdot \exp(-E/RT_{(t)})$$
  
= A \cdot exp(-E/RT) \cdot exp[(-A/\beta) \cdot fexp(-E/RT) \cdot dT] (8)  
Ln(\beta) \approx C\_1 - (E/R \cdot T) (9)

where  $\beta$  corresponds to the heating rate dT/dt (in °C.min<sup>-1</sup>) and  $C_1$  is constant. Thus the value of activation energy, E can be obtained from the slope of the plot of  $Ln(\beta)$  as a function of reciprocal Temperature (1/T). The ASTM standard method [41] for determination of Arrhenius constants relies on a variant of the Ozawa method [37]. According to the standard, an experimenter has to measure a series of kinetic curves at different heating rates. In practice, for an accurate statistical manipulation, the experiment should include measurements at four or more different heating rates [36]. In practice, two experiments with two heating rates are sufficient for these calculations but an acceptable estimation of E is only possible in the case where these heating rates considerably differ. The present kinetic parameters were not estimated from a modulated thermo gravimetric analysis (MTGA) and because of the limitations of our equipments it was not possible to perform the present kinetic studies with heating rates considerably different one to each other. However, if the kinetic studies are performed in absence of different heating rates,  $\beta$  is equal to zero and concomitantly the

 $[exp[(-A/\beta) \cdot ]exp(-E/RT) \cdot dT]$  member in equation 8 is equal to the unit, and thus, equation 8 reduce to classical Arrhenius expression from Eq. (4). It's important to improve the technical conditions of our equipment in order to show this approximation is valid to demonstrate the reliability of the present results. In addition, the present activation energy of about 16.1 kcal.mol<sup>-1</sup> (Table 6) obtained under  $N_2$  flow in the pyrolysis zone of low temperature is in good agreement with the value obtained by López-Pasquali and co-workers of about 16.7 kcal.mol<sup>-1</sup> employing a constant hating rate for the carbonization of another type of Algarroba wood [42]. Finally, it should be pointed out that carbon derived materials obtained from sawdust of wood contain lower ash content that those prepared from other lignocellulosic materials such as agroindustrial bio-wastes and clearly much lower than those obtained from petroleum precursors. This is one of the reasons why activated carbon materials prepared from sawdust of wood has been employed successfully in catalytic heterogeneous reactions [1]. For example, Laine et al. [4, 5] have showed that pore volume of activated carbon supports play an important synergistic role upon the activity and selectivity of NiMo catalysts the in thiophene hydrodesulphurization. Also, our group has showed that pore size distribution clearly influence the catalytic activity of Ni and NiMo catalysts in the ethylene hydrogenation [2] and the kinetics of coke deposition [3]. Also, our group have showed in different works about synthesis of activated carbons by physical activation or by pyrolysis [9], and by chemical activation [10] that the pore size distribution and surface area of activated carbons remarkably affects the photoactivity of  $TiO_2$  in the photocatalytic detoxification of 4-chlorophenol. In addition, we have showed that textural properties of activated carbon clearly influence the selectivity of main intermediate products detected during the aromatic molecules as phenol and 2,4-dichlorophenoxiacetic acid [6,7] and more recently on 4-chlorophenol [43,44] and 2-propanol [45] photooxidations. In this sense, the Fig. (4) shows the influence of two carbons upon the phenol adsorption and on the photoefficiency of TiO<sub>2</sub> in the phenol photo detoxification under UV-irradiation was studied as a targeted application. It can be seen in Fig. (4) that any of two TiO<sub>2</sub>-AC binary materials adsorbed higher phenol (after 15min adsorption in the dark). This enhancement in phenol molecules around photoactive  $TiO_2$  enhances the photo efficiency of the semiconductor as can also be seen in Fig. (4). This enhances has been attributed to the presence of a contact interface between TiO<sub>2</sub> and AC that make possible a continuous transfer of the species from the AC to the  $TiO_2$  surface [45].

Our present enforces are aimed to prepare hierarchically macro-meso-micro porous carbon materials to study the influence of pore size distribution on the selectivity of NiMo catalysts in hydrocracking reactions and to verify the presence of confinement effects on the selectivity of hydrocracking consequence of specific pore size an pore volume of the support such as in the case of zeolites [46,47]. We do believe



**Fig. (4).** Summary of kinetic results of phenol adsorption in the dark ( $Ph_{ads}$ ) and first-order apparent rate-constants ( $k_{app}$ ) of phenol photodegradation under UV-irradiation.

that the present results regarding the pyrolysis of sawdust of a hard wood as Algarroba consists of essentially 3 different zones is a very important finding and deserve to be studied carefully. A better explanation for the present results where the reaction rates rather decreased with the increasing temperature from 350 to 600°C could be due to the presence of different chemical structures of the starting materials at zero holding time after heating up to the set temperatures (i.e. 350, 400 and 600°C). Therefore, the estimation of kinetic parameters from data of modulated thermogravimetric analysis is required to better understand the correlation of the influence of each pyrolysis zone on the pore size distribution and pore volume of carbon materials. In addition, the influence of ashes of lignocellulosic carbon precursors and the influence of additives (chemical activators) that can play the role of catalysts to improve the textural properties are necessary to clarify the significance of the present results. In this way, our groups have already reported preliminary studies regarding the influence of the pyrolysis atmosphere [48] and the effect of chemical additives [49] on the topological organization of carbon materials obtained from the controlled pyrolysis of saccharose.

# **4. CONCLUSION**

The synthesis and textural characterization of carbon materials from the sawdust of Algarroba wood have been presented. First-order apparent constants (kap) and activation energies (E<sub>a</sub>) estimated from the synthesis of carbon materials suggest the presence of three different pyrolysis zones. Incipient carbonization between 200-350°C; then an extensive carbonization and re-arrangement of graphene layers between 350-600°C, and finally the micropore formation at temperature higher than 600°C. Phenol photodetoxification test were also performed to show the targeted application of these carbon materials. Results presented in the present work permits will permit to design and formulate potential catalysts for the selective biomass conversion as a function of the pyrolysis zone. Also, the present results showed that carbonbased photo catalysts would be able for the photooxidation of aromatic compounds.

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# NOMENCLATURE

A	=	Arrhenius frequency factor, min <sup>-1</sup>
BET	=	Brunauer-Emmet-Teller method
$E_a$	=	Energy of activation, kcal.mol <sup>-1</sup>
HJ	=	Harkins-Jura method
KH	=	Horvath-Kawazoe method
%m <sub>o</sub>	=	Mass remaining at zero reaction time
$\%m_t$	=	Mass remaining in the time t
$\mu_{porearea}$	=	Micropore area
$\mu_{porevolume}$	=	Micropore volume

$[(P_nO_{3n})]n$ -	=	Polyphosphate cyclic
$[(P_nO_{3n+1})](n+2)-$	=	Polyphosphate linear
t	=	Reaction time
Т	=	Temperature, °C
$W_{pore}$	=	Width mean pore
Subscripts		
a	=	Activation
ap,Ti	=	Components
kapp-phenol	=	Components
0	=	Zero reaction time
pore	=	Component
pore area	=	Component
porevolume	=	Component
R	=	Gas constant
t	=	Reaction time
Acronyms		
AA	=	Atomic absorption spectroscopy
AC	=	Activated carbon
$C_o$	=	Initial phenol concentration
$C_t$	=	Remaining phenol concentra- tion:
ICP-AES	=	Inductively-coupled plasma atomic emission spectropho- tometer
k <sub>ap,Ti</sub>	=	First-order apparent rate con- stant for the thermal degradation of wood
k <sub>app-phenol</sub>	=	First-order apparent rate- constants for the phenol photo detoxification

#### REFERENCES

- F. Rodríguez-Reinoso, "The role of carbon materials in heterogeneous catalysis", *Carbon*, vol. 36, pp.159-175, 1998.
- [2] J. Matos, J. Laine, and J. L. Brito, "Activated carbon supported Ni-Mo: effects of pretreatments and composition on catalyst reducibility and on ethylene conversion", *Appl. Catal. A Gen.*, vol. 152, pp. 27-42, 1997.
- [3] J. Matos, and J. Laine, "Ethylene conversion on activated carbon supported NiMo catalysts: effect of the Support", *Appl. Catal. A: Gen.*, vol. 241, pp. 25-38, 2003.
- [4] J. Laine, F. Severino, M. Labady, and J. Gallardo, "The synergistic participation of the support in sulfided Ni-Mo/C hydrodesulfurization catalysts", J. Catal., vol. 138, pp. 145-149, 1992.
- [5] J. Laine, F. Severino, and M. Labady, "Optimum Ni composition in sulfided Ni-Mo hydrodesulfurization catalysts: effect of the support", J. Catal., vol. 147, pp. 355-357, 1994.
- [6] J. Matos, J. Laine, and J.-M. Herrmann, "Synergy effect in the photocatalytic degradation of phenol on a suspended mixture of titania and activated Carbon", *Appl. Catal. B Environ.*, vol. 18, pp. 281-291, 1998.
- [7] J. Matos, J. Laine, and J.-M. Herrmann, "Effect of the type of activated carbons on the photocatalytic degradation of aqueous organic pollutants by UV-irradiated Titania". J. Catal., vol. 200, pp. 10-20, 2001.

- [8] J. Matos, J. Laine, J.-M. Herrmann, D. Uzcategui, and J. L. Brito, "Influence of activated carbon upon titania on aqueous photocatalytic consecutive runs of phenol photomineralization", *Appl. Catal. B Environ.*, vol. 70, pp. 461-469, 2007.
- [9] T. Cordero, C. Duchamp, J.-M. Chovelon, C. Ferronato, and J. Matos, "Surface nano-aggregation and photocatalytic activity of TiO<sub>2</sub> on H-type activated carbons", *Appl. Catal. B Environ.*, vol. 73, pp. 227-235, 2007.
- [10] T. Cordero, C. Duchamp, J.-M. Chovelon, C. Ferronato, and J. Matos, "Influence of L-type activated carbons on photocatalytic activity of TiO<sub>2</sub> in 4-chlorophenol photodegradation", *J. Photochem. Photobiol. A Chem.*, vol. 191, pp. 122-131, 2007.
- [11] A. Dufour, A. Celzard, V. Fierro, E. Martin, F. Broust, and A. Zoulalian, "Catalytic decomposition of methane over a wood char concurrently activated by a pyrolysis gas", *Appl. Catal. A Gen.*, vol. 346, pp. 164-173, 2008.
- [12] A. Dufour, A. Celzard, B. Ouartassi, F. Broust, V. Fierro, and A. Zoulalian, "Effect of the micropores diffusion on kinetics of CH<sub>4</sub> decomposition over a wood-derived carbon catalyst", *Appl. Catal. A Gen.*, vol. 360, pp. 120-125, 2009.
- [13] J. Matos, K. Díaz, V. García, T. C. Cordero, and J. L. Brito, "Methane transformation in presence of carbon dioxide on activated carbon supported nickel-calcium catalysts", *Catal. Lett.*, vol. 109, pp. 163-169, 2006.
- [14] K. Díaz, V. García, and J. Matos, "Activated carbon supported Ni-Ca: influence of reaction parameters on activity and stability of catalyst on methane reformation", *Fuel*, vol. 86, pp. 1337-1344, 2007.
- [15] C. Di Blasi, and C. Branca, "Kinetics of primary product formation from wood pyrolysis", *Ind. Eng. Chem. Res.*, vol. 40, pp. 5547-5556, 2001.
- [16] C. Di Blasi, and C. Branca, "Kinetics of the isothermal degradation of wood in the temperature range 528–708K", J. Anal. Appl. Pyrol., vol. 67, pp. 207-219, 2003.
- [17] G. Várhegyi, M. J. Antal, E. Jakab, and P. Szabó, "Kinetic modeling of biomass pyrolysis", *J. Anal. Appl. Pyrol.*, vol. 42, pp. 73-87, 1997.
- [18] J. L. Banyasz, S. Li, J. Lyons-Hart, and K. H. Shafer, "Gas evolution and the mechanism of cellulose pyrolysis", *Fuel*, vol. 80, pp. 1757-1763, 2001.
- [19] M. Lopez, M. Labady, and J. Laine, "Preparation of activated carbon from wood monolith", *Carbon*, vol. 34, pp. 825-827, 1996.
- [20] J. Matos, J. Laine, J.-M. Herrmann, "Association of activated carbons of different origins with titania in the photocatalytic purification of water. Use of solar energy", *Carbon*, vol. 37, pp. 1870-1872, 1999.
- [21] R. Lukens, "ASTM Standards Wood. Particle size distribution", ASTM International, Easton, D2862-70, 30, 1977.
- [22] R. Storel, "ASTM Standards Wood: Moisture content of wood", ASTM International, Easton, D2016-74, 409, 1986.
- [23] R. Storel, "ASTM Standards Wood: Ash in wood", ASTM International, Easton, D1102-84, 409, 1986.
- [24] R. Storel, "ASTM Standards Wood: Analysis of wood charcoal", Easton, D1762-84, 409, 1986.
- [25] S. Brunauer, P. H. Emmett, and E. Teller, "Adsorption of gases in multimolecular layers", J. Am. Chem. Soc., vol. 60, pp. 309-319, 1938.
- [26] W. D. Harkins, and G. Jura, "Surfaces of solids. XIII. A vapor adsorption method for the determination of the area of a solid without the assumption of a molecular area, and the areas occupied by nitrogen and other molecules on the surface of a solid", *J. Am. Chem. Soc.*, vol. 66, pp. 1366-1373, 1944.
- [27] G. Horvath, and K. Kawazoe, "Methods for the calculation of effective pore size distribution in molecular sieve carbon", J. Chem. Eng. Jpn., vol. 16, pp. 470-475, 1983.
- [28] K. Denbigh, "The Principles of Chemical Equilibrium", Cambridge University Press, UK, pp. 449-450, 1970.

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- [29] M. Sefain, S. El-Kalyoubi, and N. Shukry, "Thermal behavior of holo- and hemicellulose obtained from rice straw and bagasse", J. Polyme. Sci. Polym. Chem., vol. 23, pp. 1569-1577, 1985.
- [30] T. Wigmans, "Industrial aspects of production and use of activated carbons", *Carbon*, vol. 27, pp. 13-22, 1989.
- [31] J. Matos, M. Labady, A. Albornoz, J. Laine, and J. L. Brito, "Topological organization and textural changes of carbon macronetworks submitted to activation with N<sub>2</sub> and CO<sub>2</sub>", *J. Mater. Sci.*, vol. 39, pp. 3705-3716, 2004.
- [32] J. Laine, A. Calafat, and M. Labady, "Preparation and characterization of activated carbons from coconut shell impregnated with phosphoric acid", *Carbon*, vol. 27, pp. 191-195, 1989.
- [33] F. E. Huggins, G. P. Huffman, N. Shah, F. W. Lytle, R. B. Greegor, and R. G. Jenkins, "In situ XAFS studies of catalyzed pyrolysis and gasification reactions of lignite", *Phys. B Condensed Mater.*, vol. 158, pp. 178-179, 1989.
- [34] S. G. Chen, and R. T. Yang, "Unified mechanism of alkali and alkaline earth catalyzed gasification reactions of carbon by CO<sub>2</sub> and H<sub>2</sub>O", *Energy Fuels*, vol. 11, pp. 421-427, 1997.
- [35] F. Rodriguez, F. Ruette, and J. Laine, "Molecular modeling of micropores in activated carbon", *Carbon*, vol. 32, pp. 1536-1537, 1994.
- [36] M. K. Bahng, C. Mukarakate, D. J. Robichaud, and M. R. Nimlos, "Current technologies for analysis of biomass thermochemical processing: a review", *Anal. Chim. Acta*, vol. 651, pp. 117-138, 2009.
- [37] T. Ozawa, "A new method of analyzing thermogravimetric data", Bull. Chem. Soc. Jpn., vol. 38, pp. 1881-1886, 1965.
- [38] V. Mamleev, S. Bourbigot, M. Le Bras, and J. Lefebvre, "Three model-Free methods for calculation of activation energy in TG", J. *Therm. Anal. Calor.*, vol. 78, pp. 1009-1027, 2004.
- [39] S. Hu, A. Jess, and M. H. Xu, "Kinetic study of chinese biomass slow pyrolysis: comparison of different kinetic models", *Fuel*, vol. 86, pp. 2778-2788, 2007.
- [40] J. J. M. Orfao, F. J. A. Antunes, and J. L. Figueiredo, "Pyrolysis kinetics of lignocellulosic materials-three independent reactions model", *Fuel*, vol. 78, pp. 349-358, 1999.
- [41] "ASTM standard test method for decomposition kinetics by thermogravimetry", ASTM International, West Conshokocken, E1641, 2007.
- [42] C. E. López-Pasquali, C. A. Wottitz, R. G. Martínez, and H. A Herrera, "Carbonization of Algarrobo Negro (*Prosopis Nigra*): a study of its microstructure and main volatile components", *Latin. Amer. Appl. Res.*, vol. 32, pp. 321-325, 2002.
- [43] J. Matos, A. García, T. Cordero, J.-M Chovelon, and C. Ferronato, "Eco-friendly TiO<sub>2</sub>-AC photocatalyst for the selective photooxidation of 4-chlorophenol", *Catal. Lett.*, vol. 130, pp. 568-574, 2009.
- [44] J. Matos, A. García, and P. S. Poon, "Environmental green chemistry applications of nanoporous carbons", J. Mater. Sci., vol. 45, pp. 4934-4944, 2010.
- [45] J. Matos, E. García-López, L. Palmisano, A. García, and G. Marci, "Influence of activated carbon in TiO<sub>2</sub> and ZnO mediated photoassisted degradation of 2-propanol in gas-solid regime", *Appl. Catal. B Environ.*, vol. 99, pp. 170-180, 2010.
- [46] E. Benazzi, L. Leite, N. Marchal-George, H. Toulhoat, and P. Raybaud, "New insights into parameters controlling the selectivity in hydrocracking reactions", *J. Catal.*, vol. 217, pp. 376-387, 2003.
- [47] H. Toulhoat, P. Raybaud, and E. Benazzi, "Effect of confinement on the selectivity of hydrocracking", J. Catal., vol. 221, pp. 500-509, 2004.
- [48] J. Matos, and J. Laine, "A carbon macro-network from the controlled pyrolysis of sacharose", J. Mater. Sci. Lett., vol. 17, pp. 649-651, 1998.
- [49] J. Matos, M. Labady, A. Albornoz, J. Laine, and J. L. Brito, "Catalytic effect of KOH on textural changes of carbon macronetworks by physical activation", *J. Mol. Catal. A Chem.*, vol. 228, pp. 189-194, 2005.

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