

Characteristics and Microbial Succession in Co-Composting of Oil Palm Empty Fruit Bunch and Partially Treated Palm Oil Mill Effluent

Azhari Samsu Baharuddin^{1,2}, Nakamura Kazunori⁴, Suraini Abd-Aziz^{*-3}, Meisam Tabatabaei³, Nor' Aini Abdul Rahman³, Mohd Ali Hassan³, Minato Wakisaka¹, Kenji Sakai⁴ and Yoshihito Shirai¹

¹Department of Biological Functions and Engineering, Graduate School of Life Science and System Engineering, Kyushu Institute of Technology, 2-4 Hibikino, Wakamatsu-ku, Kitakyushu, Fukuoka 808-0916, Japan

²Department of Process and Food Engineering, Faculty of Engineering, Universiti Putra Malaysia, 43400 UPM Serdang, Selangor, Malaysia

³Department of Bioprocess Technology, Faculty of Biotechnology and Biomolecular Sciences, Universiti Putra Malaysia, 43400 UPM Serdang, Selangor, Malaysia

⁴Department of Plant Resources, Graduate School of Bioresources and Bioenvironmental Sciences, Kyushu University, 6-10-1, Higashi-ku, Fukuoka, 812-8581, Japan

Abstract: Microbial communities and cellulolytic enzymes activities were analyzed during the co-composting of empty fruit bunch (EFB) and partially treated palm oil mill effluent (POME) in pilot scale. The physicochemical parameters were also measured during the composting. The diversity of the bacterial community was investigated using polymerase chain reaction-denaturing gradient gel electrophoresis (PCR-DGGE). The results indicated that the composting process of EFB with partially treated POME was dominated by uncultured bacteria species. The dominant bacterial group changed from the phylum *proteobacteria* in the thermophilic stage to the phylum *chloroflexi* in the maturing stage. Scanning electron microscope (SEM) analysis exhibited the significant degradation of EFB structure during the composting process. The maximum cellulase activity for carboxymethylcellulase (CMCase), filter paperase (FPase) and β -glucosidase were 13.6, 4.1 and 20.3 U/g of dry substrate, respectively at day 30 of composting. The results of this study significantly contributed to a better understanding of mechanisms involved in co-composting process in pilot scale.

Keywords: Composting, empty fruit bunch, partial treated palm oil mill effluent, cellulase, DGGE.

INTRODUCTION

Oil palm production is a major agricultural industry in Malaysia. The oil palm biomass (OPB) produces about 40 million tonnes per year. This OPB can be categorized as a form of empty fruit bunches (EFB), oil palm trunks (OPT) and oil palm fronds (OPF) and the rest are palm oil mill effluent (POME). Composting has been considered to be one of the most suitable ways of converting organic wastes into products that are beneficial for plant growth. Thus, the utilization of EFB and POME as substrate for the composting is crucial to convert into value added product and treat the biomass simultaneously.

In our previous work, it showed that co-composting of EFB and partially treated POME can be used as an alternative method for converting of these materials using a simple and traditional windrow composting piles in pilot scale [1]. Although considerable research on composting of various organic wastes has been conducted [2,3], little information on characteristics and microbial changes from EFB and

POME is available especially in pilot scale operation. Composting is basically a biodegradation process in which organic wastes are transformed and stabilized by the metabolic activities of a succession of mixed microbial population. The enzymatic process and pre-dominant microbes are specific to the materials to be composted and need to be determined by specific experiments. Thus, the information would provide specific design and operational parameters for effective composting of EFB and POME.

The decomposition of cellulose is mediated by group of enzymes collectively called as cellulases. The decomposition of organic waste during the composting process is carried out by a succession of microbial communities, which is critical for the utilization of complex substrates such as cellulose, hemicellulose and lignin [4]. DGGE is a useful tool to reveal microbial succession during composting. Many researchers showed the efficiency of this molecular method to detect the shift of microbial changes in composting process [5-7]. Therefore, the aim of this study was mainly focused on the observation and identification of enteric microorganisms, biochemical changes and cellulase profiles during the co-composting of EFB with partially treated POME in pilot scale. DGGE technique was used to characterize microbial communities and diversity during the composting process.

*Address correspondence to this author at the Department of Bioprocess Technology, Faculty of Biotechnology and Biomolecular Sciences, Universiti Putra Malaysia, 43400 Serdang, Selangor, Malaysia; Tel: +6-03-89467479; Fax: +6-03-89467510; E-mail: suraini@biotech.upm.edu.my

MATERIALS AND METHODS

Pilot-Scale Composting Plant

This research was conducted at pilot-scale composting plant in FELDA Maokil, Johor, Malaysia. A total of 202 windrows were incubated at the composting site which consists of two field area (13 hectare). The composting materials were obtained from the oil palm processing mill. The shredded EFB and partially treated POME were mixed at 40 tonnes and 120 tonnes, respectively.

Composting Establishment

The tipper lorry was used to lay down the shredded EFB on the composting ground to form a long conical shaped of windrow (40m L x 3m W x 1.5m H). Partially treated POME (anaerobic pond) was added to each of the piles every three days interval to adjust the final moisture content of 65-70%. The addition of POME was stopped one to two weeks prior to harvesting to avoid the final product from being too wet. The windrow was covered throughout the composting period except during spraying of POME and turning process. The piles were turned over one to three times per week to maintain an even distribution of moisture and prevent the build-up of heat.

Sampling

A 2 kg sample was collected at different locations of the windrow: bottom, core and surface. The samples were divided into two parts. One part was stored at 4°C while the other part was air dried, and then passed through 2 mm sieve. All experiments were done in triplicates.

Physicochemical and Bacterial Count Analysis

The temperature at the top, middle and bottom layers of the interior of the heaps was monitored by using compost monitoring device for temperature and oxygen. pH meter and moisture analyzer were used in this study for pH and moisture content analysis. Total carbon and total nitrogen were measured using a CNHS analyzer. Plate count method was used to obtain viable bacterial count. Observation of the compost structure was carried out using scanning electron microscopy (SEM).

Extraction Procedures for Cellulases Analysis

The mixed compost sample (5 g) was extracted with 50 mL of 0.01 M phosphate buffer (pH 7) at 4°C. Then, the sample was centrifuged at 5000 rpm, 4°C for 20 minutes and the supernatant was stored at -20°C. The supernatant obtained was then used for the analysis of cellulases activities [8].

Enzyme Assays and Reducing Sugar Concentration

Enzyme assays for CMCase were performed by measuring reducing sugars from 2% carboxymethylcellulose (CMC). Filter paperase (FPase) activity was determined by estimating the reducing sugars from Whatman filter paper no 1. The β -glucosidase activity was determined using p-nitrophenol- β -D-glucopyranoside as substrate. The reducing sugars released were analyzed according to the dinitrosalicylate (DNS) method. One unit (U) of enzyme activity was defined as the amount of enzyme required to liberate 1 μ mol

of product per min. Cellulases enzyme assays (CMC, FPase, β -glucosidase) and reducing sugars determination was followed the methods reported by Shahrim *et al.*, 2008 [8]. Enzyme yield was expressed as U/g of dry substrate [8].

Nucleic Acid Extraction

Total genomic DNAs were extracted from approximately 1.0 g of mixed compost samples using Ultra Clean Soil DNA Isolation Kit (MoBio laboratories, USA). The DNA extracted was checked for size and quality. Extracted DNA samples were then purified using DNA purification kit with GeneCleanR Turbo kit (MP Biomedicals, USA).

Polymerase Chain Reaction

The variable region (V3 region) of the 16S rDNA was amplified by using PCR primers. The PCR amplifications were performed in 50 μ l volumes containing approximately 1 μ l template DNA, 25 μ l Ex Taq DNA polymerase (Takara Shuzo, Japan), 20 μ l ultra pure water (Millipore) and 2 μ l of each primer: 341F-with a 40-bp GC clamp (5'-CGC CCG CCG CGC GCG GCG GGC GGG GCG GGG GCA CGG GGG GCC TAC GGG AGG CAG CAG -3') and 518R (5'-ATT ACC GCG GCT GCT GG-3'). Combination of these primers generates a PCR fragment of about 200 bp. The PCR cycling was performed using a Takara PCR Thermal Cycler. The temperature used was as follows; 94°C for 3 min followed by 30 cycles of 52°C for 1 min, 72°C for 1 min, 94°C for 1 min and then continue at 52°C for 1 min with final extension steps at 72°C for 10 min.

Denaturing Gradient Gel Electrophoresis

DGGE was performed with a D-Gene system (BioRad laboratories, USA). PCR products were separated on a 1.5mm thick vertical gel containing 8% (w/v) acrylamide and a linear gradient of the denaturants urea and formamide, increasing from 30% at the top of the gel to 70% at the bottom. One hundred percent of denaturant corresponds to 7M urea and 40% (v/v) formamide. The PCR products (200ng DNA) were then applied to the individual lanes in the gel. Electrophoresis was performed in a TAE buffer and a constant voltage of 200 V was applied to the submerged gel for 5h at 60°C. After electrophoresis, the gels were stained with SYBR^R Green nucleic Acid Gel Stain for 30min and then rinsed with water and photographed on a UV transillumination table. The computerized images were then inverted to negative images. Small pieces of selected DGGE bands viewed on the UV were excised from the gel with pasteur pipettes and eluted in 50 μ l TE buffer for overnight. The DNA fragments in the gels were washed and directly re-amplified with the same primer. The PCR products were confirmed by DGGE as a single band and then sequenced. Before sequencing, the PCR products were purified by Qiaquick PCR Purification kit (QIAGEN, Hilden, Germany).

Sequencing and Band Characterization

The PCR products were sent for sequencing. Sequence similarity searches were conducted using the BLAST (Basic Local Alignment Search Tool) network service of the GenBank database through website (<http://www.ncbi.nlm.nih.gov/>) and Ribosomal Database Project II (RDP) to identify the nearest relatives of the partially sequenced 16S rRNA genes and the excised dominant bands.

Table 1. Characteristics of Raw Materials and Compost Samples for the Denaturant Gradient Gel Electrophoresis (DGGE) Analysis

Sample	Color	Temperature (°C)	pH	Moisture Content (%)	C/N	Plate Count (CFU/g x 10 ¹⁰)
EFB	Brown	32.3	6.5	25	56.5	-
POME	Blackish	41.3	7.6	95	13.6	-
Day 2	Brown	60.2	8.0	70	44.5	4.4±2.5
Day 10	Brown	60.0	8.5	68	38.5	20.8±3.3
Day 20	Dark brown	58.6	8.5	67	28.5	18.3±0.5
Day 30	Dark brown	57.4	8.3	65	24.6	17.6±1.7
Day 40	Blackish	48.1	8.2	65	20.2	14.1±2.1
Day 45	Blackish	43.3	8.1	62	15.7	12.2±0.9
Day 60	Blackish	35.3	7.8	60	12.8	11.0±1.8

RESULTS AND DISCUSSION

Physicochemical Characteristics

The changes in temperature, moisture content, pH, C/N and colour throughout composting process were monitored as shown in Table 1. The final matured compost was found blackish in color and has an earthy smell. Based on temperature profile, two compost phases were involved; thermophilic and curing phase (Table 1). The temperature in the compost piles reached about 60°C after second day of composting and fluctuated in the range of 50-60°C until day 40 and gradually descended to ambient values thereafter. This result might be due to the large amount of readily digestible components contained in the POME added which were immediately available for microbes. It was observed that the moisture content throughout the treatment was 65-70% and gradually decreased to 60% for final compost. The higher moisture content (65%) enhanced the microbial activities in the composting process, resulting in higher biodegradation of organic compounds [9].

The pH profile increased gradually from 7.8 to about 8.0 on the second day treatment and remained at 8.2-8.5 as the temperature increased in the thermophilic phase indicating weak alkaline condition of the system (Table 1). The pH was then gradually decreased afterwards and reached about 7.8 in the final compost. The initial C/N ratio of the piles was about 44.5 and decreased to 12.8 (Table 1) after 60 days of treatment. In the latter stage of treatment, the number of bacteria decreased when most of the easily metabolisable carbon constituents have already been exhausted.

SEM Analysis and Change in Cellulolytic Activities

The structure of EFB before and during composting treatment was viewed under SEM as shown in Figs. (1 to 3). SEM observation showed that the surface structure of the untreated shredded EFB consisted of firmly bound threads of lignin with smooth surface along the structure (Fig. 1a) while the outer surface of treated EFB seems has been altered with the presence of many holes indicating the lignin has been disrupted (Fig. 3).

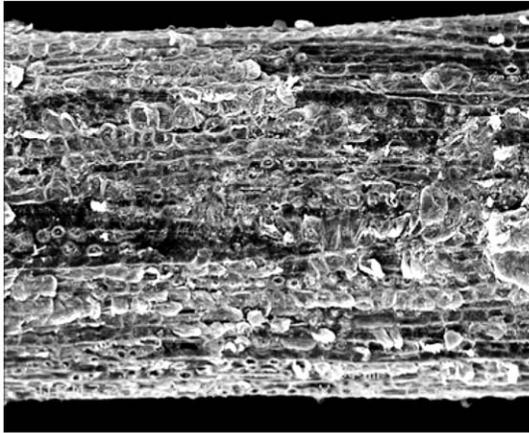
Moreover, there were many silica bodies found on the strand surface of the EFB structure at initial composting

process (Fig. 2). This result was in agreement with previous data published by Law *et al.*, 2007 [10], who reported that the silica bodies were spread uniformly over the EFB strand's surface. Interestingly, most of the silica bodies on the EFB strand were removed when the composting process achieved at day 20 of treatment (Fig. 3). After day 40 of treatment, the inner structure of EFB compost was strongly disrupted. This phenomenon might be due to the combination of thermophilic temperature (>50°C) which occurred from day 3 to day 40 of treatment, microbial degradation and the frequent turning on the composting piles. The removal of silica bodies would enhance the microbial penetration in the composting process.

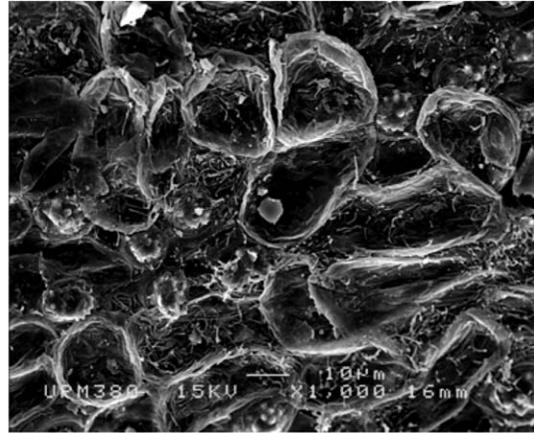
The changes of CMCase, FPase and beta-glucosidase were measured (Fig. 4). After day 3 of treatment, when the temperature increased into thermophilic phase (>50°C), the cellulase activity also increased rapidly. During the thermophilic phase, high temperatures accelerated the breakdown of proteins, fats and complex carbohydrates such as cellulose and hemicellulose [11]. The highest activity of CMCase, FPase and β -glucosidase was detected at day 30 of treatment with values of 13.6, 4.1 and 20.3 U/g of dry substrate, respectively. During the curing phase, C/N was less than 15 (day 45) while the cellulase activity remained unchanged until the end of the process.

CMCase activity started to increase until day 10 and showed the maximum activity at day 30 before it slightly declined afterwards. The shredded EFB used in this study may result in an enlargement of inner surface area of substrate particles, and was accomplished by partial solubilization or degradation of hemicellulose and lignin. The activity of FPase corresponded to the degradation of cellulosic material from EFB after the action of CMCase at the early stage. Since CMCase is not very active against crystalline cellulose, FPase will take over the action to degrade the insoluble cellulose. There is a synergistic effect between CMCase and FPase. These two enzymes must be incorporated towards each other so that complete hydrolysis of cellulose was achieved. The highest amount of secreted β -glucosidase was at day 30 (20.3 U/g) and slightly decreased afterwards. The activity of β -glucosidase was correlated to the amount of cellobiose produced from the action of FPase. Based on the

a) Shredded EFB

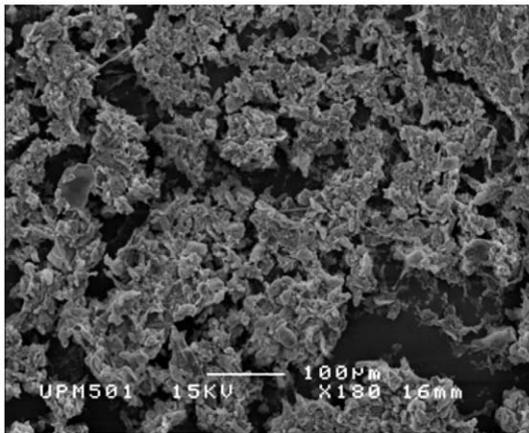


(180X magnification)

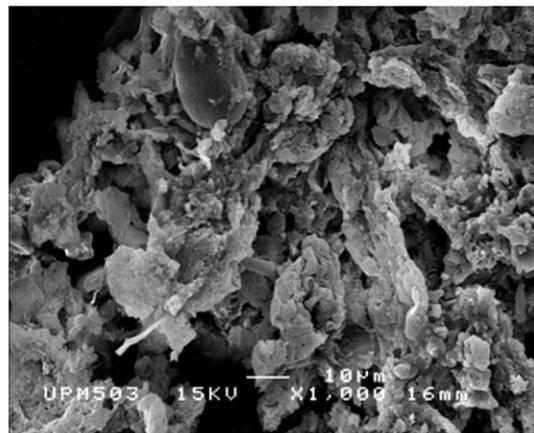


(1000X magnification)

b) Partially treated POME



(180X magnification)



(1000X magnification)

Fig. (1). SEM of shredded EFB and partially treated POME.

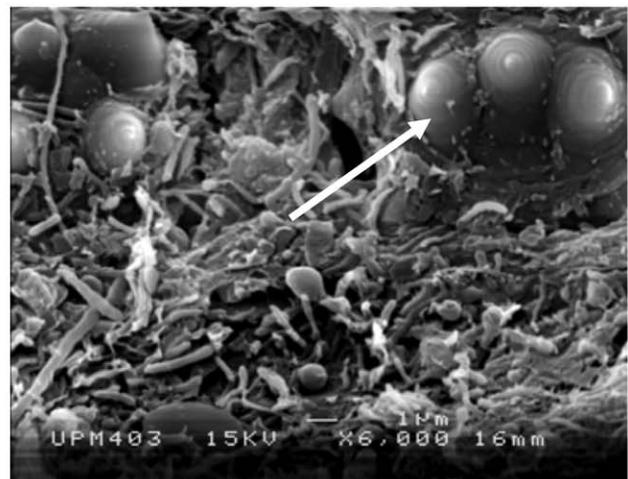
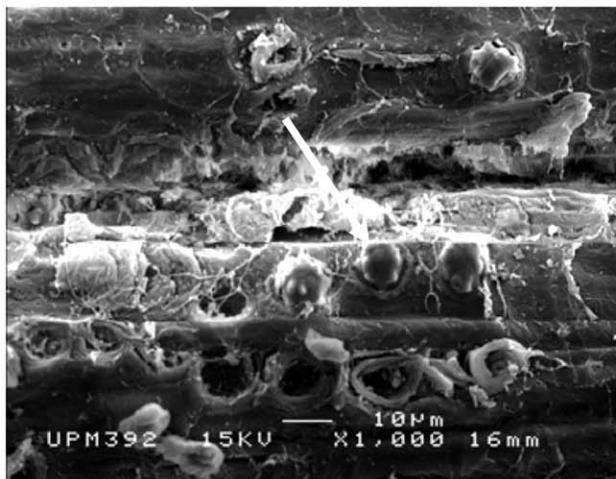


Fig. (2). SEM of EFB compost at day 2 (arrow shows the silica body found on the EFB strand at 1000 and 6000X magnification).

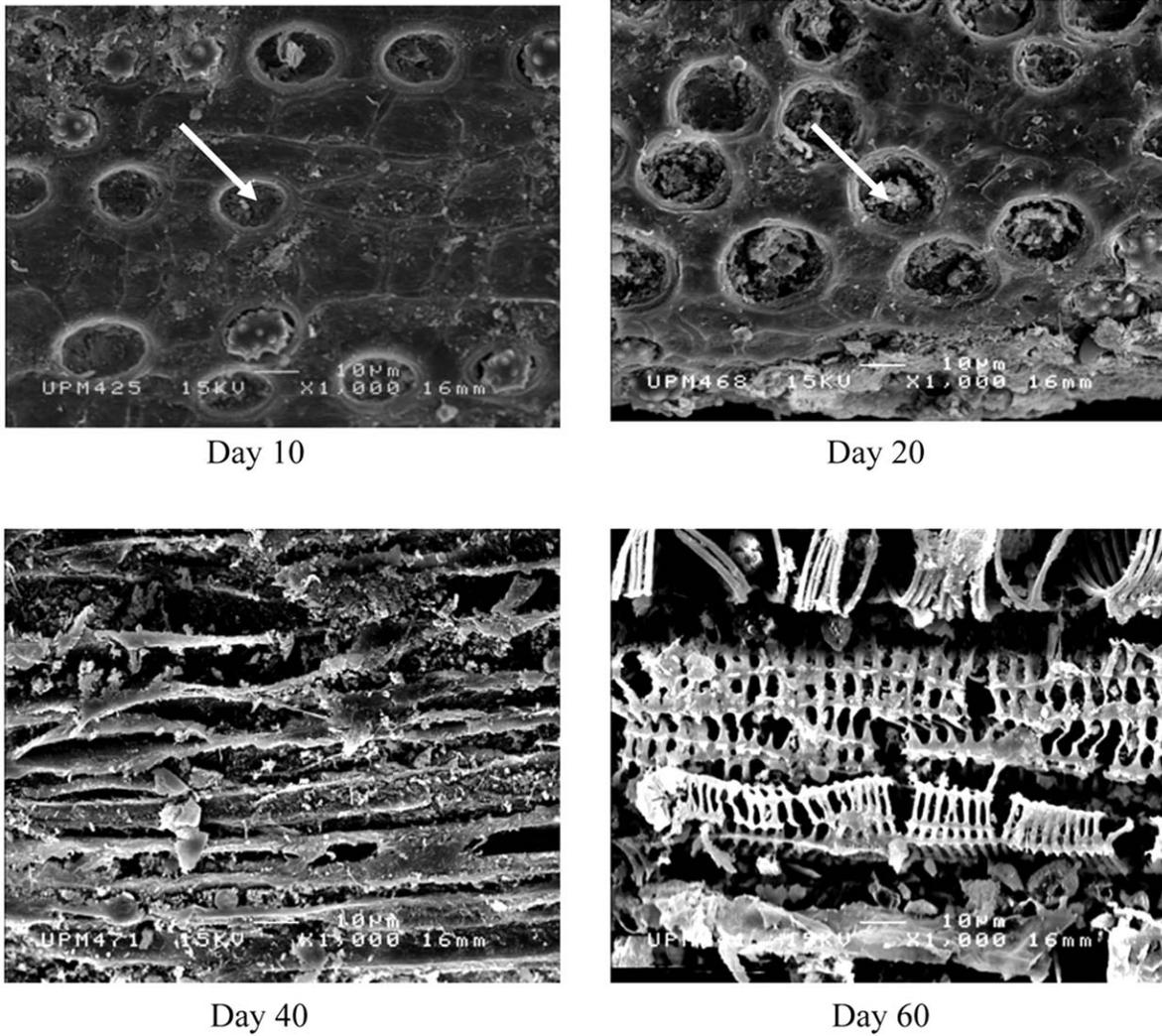


Fig. (3). SEM of EFB compost at day 10, day 20, day 40 and day 60 at 1000X magnification (arrow shows the silica body found on the EFB strand).

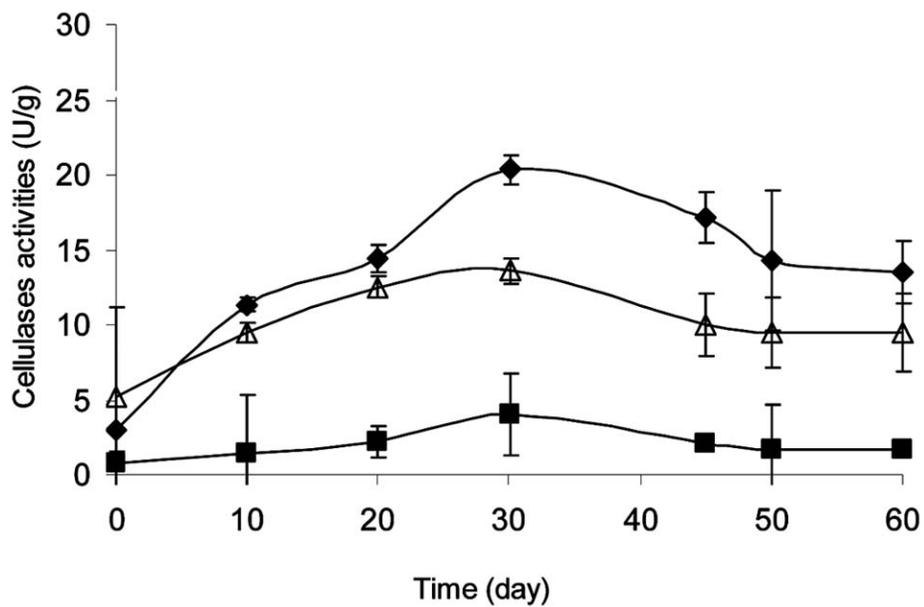


Fig. (4). Change in cellulase activities (U/g dry substrate) during the co-composting of EFB and treated POME. Values are means of three replicates \pm standard errors. (\blacksquare : FPase; Δ : CMCCase; \blacklozenge : β -glucosidase).

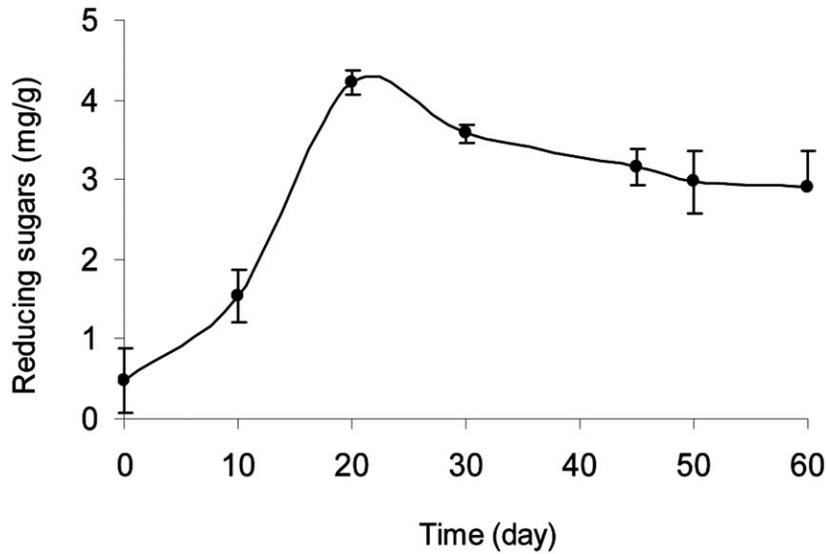


Fig. (5). Profiles of reducing sugar production (mg/g dry substrate) throughout composting.

high activity of cellulases produced during composting indicated the degradation of EFB and were in agreement with the SEM results. The profile of reducing sugar production is shown in Fig. (5). The highest reducing sugar produced was detected at day 20 (4.2 mg/g of dry substrate). The decreased of reducing sugars at day 30 might be due to the sugar consumption by the microorganism. The reducing sugars continued to decrease and remained stable during curing phase (C/N: 15.8) after 45 days of treatment.

DGGE Analysis in Composting Process

In this study, composting process in pilot scale was analyzed in terms of biochemical aspects and bacterial populations evaluated by PCR-DGGE analysis. The results showed that the position and intensities of most bands did not change significantly as the age of composting proceeded (Fig. 6). These findings suggest that bacterial communities did not change greatly throughout the composting treatment. However, the detailed analyses of each sample showed the pres-

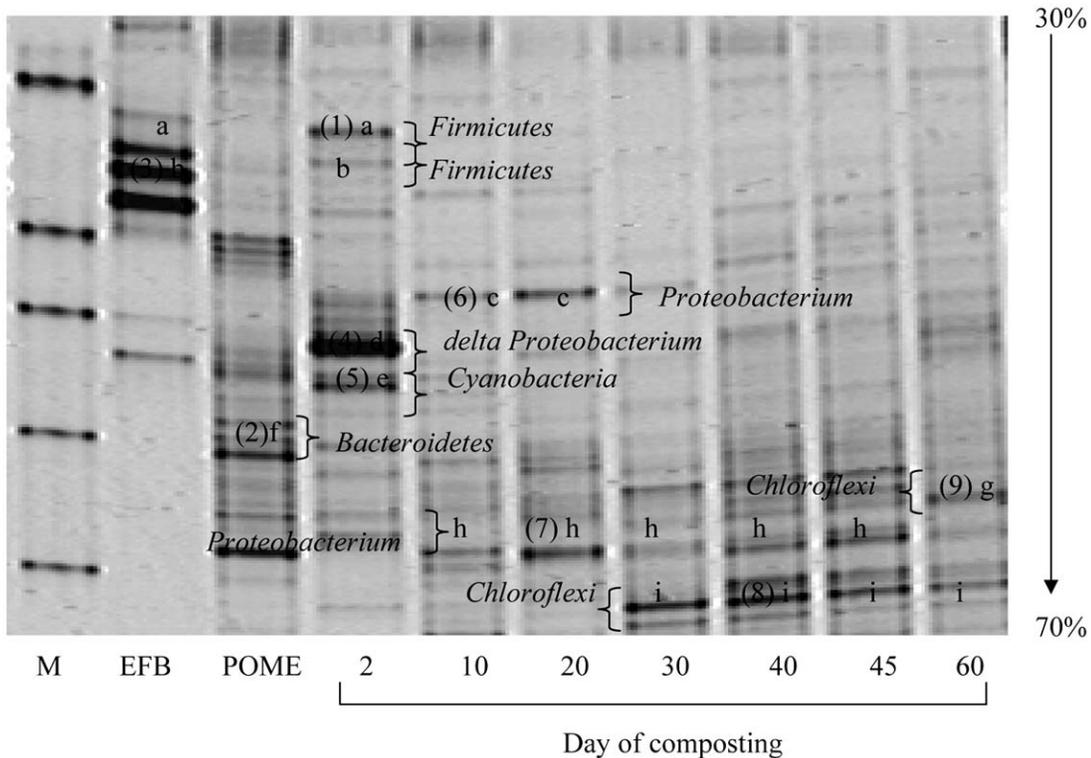


Fig. (6). Comparison of the DGGE banding patterns of microbial communities of the substrates at various composting time. The arrow on the right indicates the gradient of DNA denaturants.

Table 2. Phylogenetic Affiliation of Excised DGGE Bands

Band	Nearest Relative (Accession)	Similarity (%)	Source	Phylogeny	Nearest Known Relatives (Accession)	Similarity (%)	Source
1	Uncultured bacterium clone:C21-397 (AB366968.1)	94	EFB	<i>Firmicutes</i> "Bacili"	<i>Lactobacillus agilis</i> , strain: PIL54 (AB300541.1)	94	Solid waste compost
2	Uncultured biogas bacterium clone ATB-KS-1413 (EF686945.1)	96	POME	<i>Bacteroidetes</i>	Uncultured <i>Bacteroidetes</i> bacterium clone 1217 (EF109915)	91	High-temperature volcanic
3	Uncultured bacterium clone:C21-397 (AB366968.1)	94	compost	<i>Firmicutes</i> "Bacili"	<i>Lactobacillus</i> sp. <i>Autruche</i> 4 (DQ418552.1)	92	Probiotics on chicks
4	Uncultured bacterium clone 31b02 (EF515618.1)	90	compost	<i>Proteobacterium</i> "δ-proteobacterium"	Uncultured <i>Desulfobacteraceae</i> bacterium clone cLaKi-StA5 (AJ582696.1)	90	Full-scale anaerobic bioreactor sludge
5	Uncultured bacterium C051 (AF128638.1)	94	compost	<i>Cyanobacteria</i>	Uncultured <i>Oscillatoriales cyanobacterium</i> (AY795633.2)	94	Soil
6	Uncultured bacterium clone biogas-DT-B19 (DQ419742.1)	86	compost	Bacteria	Uncultured <i>proteobacterium</i> clone GASP-KA1W1_C01 (EU297196.1)	86	Crop land
7	Uncultured compost bacterium,clone G818 (AJ011365.1)	91	compost	Bacteria	Uncultured <i>Proteobacterium</i> isolate DGGE band CSMC-M (EU025084.1)	91	Hot Compost
8	Uncultured bacterium clone DGGE band C-2 (DQ082874.1)	96	compost	Bacteria	Uncultured <i>Chloroflexi</i> bacterium clone GASP-KB1S2_D03 (EU297607.1)	96	Soil
9	Uncultured Chloroflexi bacterium clone MSB-5B12 (DQ811882.1)	93	compost	<i>Chloroflexi</i>	Uncultured <i>Chloroflexi</i> bacterium clone MSB-5A11 (DQ811881.1)	93	Soil

ence of distinguishable patterns that revealed the different bacterial genera or species in each community (Table 2). Moreover, the bacteria involved in the composting process might not totally derived from the starting materials, because the DGGE profiles of compost samples were different as compared to the band detected in the raw material (Fig. 6). It was shown that different microbial populations from those in the raw material rapidly developed in the composting process. The result was supported by kramer and Baath, 1998 [12] who reported that the microbes proliferating in the composting processes would be adapted to the composting environment and selected by factors within the composting materials.

The results obtained in this study indicated that the prominent members were mainly uncultured and unidentified bacteria. This result is in agreement with previous studies [13] on the identification of thermophilic bacteria in EFB

compost. The recovered sequences from major bands were distributed unevenly among 6 phylogenetic groupings: *Firmicutes* (2 sequences), *Bacteroidetes* (1 sequence), *Cyanobacteria* (1 sequence), *delta-Proteobacterium* (1 sequence), *Proteobacterium* (2 sequence), and *Chloroflexi* (2 sequences) (Fig. 6). The majority of these bacteria had high similarity to sequences of known bacteria or environmental sequences (8 sequences, equal to or greater than 90% identity), although one sequence, uncultured bacterium clone biogas-DT-B19 (*proteobacterium*) for sample of day 10 and 20, was more distantly related to known sequences (<90% identity) (Table 2).

In this study, the phylum of *Firmicutes*, the closest to *Lactobacillus*, was the only prominent detected in EFB (Table 2). This result could be attributed to the remaining oil content in the EFB. It can be seen that the pH of the shredded EFB was slight in acidic condition (6.5). According to

Hemmi *et al.*, 2004 [14], in the presence of oil, lactic acid bacteria become the major strain of microflora. The freshly sterilized empty fruit bunches usually still contained about 2% of crude palm oil [15]. The uncultured bacterium clone C21-397 (AB366968.1) detected in this study has been found during start-up period of large scale solid waste compost as reported by Watanabe *et al.*, 2007 [16], while in POME (anaerobic pond) it was dominated by the members of the phylum *bacteroidetes* (uncultured biogas bacterium) (Table 2).

At the beginning of composting, when the POME was added onto EFB and turned, sample of day two was taken and analyzed through DGGE analysis. Three major bands were excised and known as a group of *cyanobacteria*, *delta-proteobacterium* and *firmicutes* (Fig. 6). The phylum of *firmicutes* existed at this time may be due to the prominent bacterial detected in EFB (oil). Once the temperature of the pile increased to 44°C (thermo-tolerant), the temperature encouraged other predominant bacterial presence, namely *cyanobacteria* and *delta-proteobacterium* group. The closest known relative to the excised band number 4d (*delta-proteobacterium*) was uncultured *desulfobacteraceae*. The uncultured bacterium detected for the excised band number 5e (*cyanobacteria*) was closest to *oscillatoriales cyanobacterium*.

Major band number 6c which was detected at day 10 and 20 was identified as uncultured bacterium clone biogas and closest to the phyla *proteobacterium* bacteria. During this time, the composting was still unstable and this bacterium could be obtained from the POME added to the piles, while the major bands of 7h, which were consistently present from day 10 to day 45, were affiliated to uncultured compost bacterium. The closest known relative of this bacterium was *proteobacterium* and was previously isolated from hot compost sample [17]. However, once the composting condition achieved C/N about 25 (pre-matured compost) at day 35, the uncultured bacterium closest to *chloroflexi* bacterium was dominated in the compost together with *proteobacterium* (Fig. 6). After day 45 (C/N ratio: 15), only uncultured *chloroflexi* bacterium was detected. The phyla *chloroflexi* are known to contain bacteria that are highly abundant in the environment, even though most of them are difficult to cultivate [18]. *Chloroflexi* and *proteobacteria* were also detected as prominent microbial community in soil [19].

The sequences of the amplified DNA bands in this DGGE study also confirmed the stability of the microflora. For example band 7h (uncultured compost bacterium) remained as the major microorganisms for over 6 weeks. The results also revealed that all of the pre-dominant sequences were most closely related to several species of uncultured bacteria found in different compost samples (Table 2). The results obtained in this study suggested that these bacteria play important roles in the co-composting of EFB and partially treated POME in pilot scale.

CONCLUSIONS

The adjustment of composting conditions such as pile size, aeration, moisture content, pH and temperature is crucial. This would increase microbial populations and their hydrolytic enzyme activities. The organic matter decomposition was increased as shown by the reduction of C/N ratio.

The results of DGGE analysis indicated that the prominent microbes during the composting of EFB and partially treated POME were mainly uncultured and unidentified bacteria.

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REFERENCES

- [1] Baharuddin AS, Wakisaka M, Shirai Y, Abd-Aziz S, Abdul Rahman NA, Hassan MA. Co-composting of empty fruit bunches and partially treated palm oil mill effluents in pilot scale. *Int J Agric Res* 2009; 4(2): 69-78
- [2] Gea T, Artola A, Sort X, Sanchez A. 2005. Composting of residuals produced in the Catalan wine industry. *Compost Sci Util* 2005; 13: 168-174.
- [3] Virginia SA, Francisco JF, Jose V, Lourdes R. Enhancing the co-composting of olive mill wastes and sewage sludge by the addition of an industrial waste. *Bioresour Technol* 2008; 99: 6346-53.
- [4] Hiroaki T, Kodaira S, Kimoto A, Nashimoto M, Takagi M. Microbial communities in the garbage composting with rice hull as an amendment revealed by culture-dependent and - independent approaches. *J Biosci Bioeng* 2006; 101 (1): 42-50.
- [5] Haruta S, Kondo M, Nakamura K, *et al.* Microbial community changes during organic solid waste treatment analyzed by double denaturing gradient gel electrophoresis and fluorescence in situ hybridization. *Appl Microbiol Biotechnol* 2002; 60: 224-231.
- [6] Ishii K, Fukui M, Takii S. Microbial succession during a composting process as evaluated by denaturing gradient gel electrophoresis analysis. *J Appl Microbiol* 2000; 89: 768-777.
- [7] Tanahashi T, Murase J, Matsuya K, Hayashi M, Kimura M, Asakawa S. Bacterial communities responsible for the decomposition of rice straw compost in a Japanese rice paddy field estimated by DGGE analysis of amplified 16S rDNA and 16S rRNA fragments. *Soil Sci Plant Nutr* 2005; 51: 351-360.
- [8] Shahrim Z, Sabaratnam V, Rahman NAA, Abd-Aziz S, Hassan MA, Karim MIA. *Res J Microbiol* 2008; 1-11.
- [9] Xiujin L, Zhang R, Pang Y. Characteristics of dairy manure composting with rice straw. *Biores Technol* 2008; 99: 359-367.
- [10] Law NL, Daud WRW, Ghazali A. Morphological and chemical nature of fiber strands of oil palm empty fruit bunch (OPEFB). *Bioresource* 2007; 2(3): 351-360.
- [11] Olfa F, Jedidi N, Hassen A. Behaviour of main microbiological parameters and of enteric microorganisms during the composting of municipal solid wastes and sewage sludge in a semi-industrial composting plant. *Am J Environ Sci* 2008; 4 (8): 103-110.
- [12] Klammer M, Baath E. Microbial community dynamics during composting of straw material studied using phospholipid fatty acid analysis. *FEMS Microbiol Ecol* 1998; 27: 9-20.
- [13] Goh CM, Liew PWY, Jong BC, Thong KL. Bacterial diversity of decomposing oil palm empty fruit bunches based on PCR and denaturing gradient gel electrophoresis of 16S rRNA gene fragments. 29th Symposium of the Malaysian Society of Microbiology. Terengganu: 24th-26th November 2007.
- [14] Hemmi H, Shimoyama T, Nakayama T, Hoshi K, Nishino T. Molecular biological analysis of microflora in a garbage treatment process under thermoacidophilic conditions. *J Biosci Bioeng* 2004; 97(2): 119-126.
- [15] Mohamad H, Ridzuan R, Anis M, *et al.* Research on development of oil palm biomass utilisation in wood-based industries. *Malaysian Palm Oil Board* 2002; 2: p. 1-5.
- [16] Watanabe K, Nagao N, Toda T, Kurosawa N. Bacterial community succession during a start-up period of large-scale composting reac-

- tor as evaluated by DGGE and clone analysis 2007. Available from:<http://www.ncbi.nlm.nih.gov/> [Cited 2008 Oct 10].
- [17] Blanc MP, Marilley L, Beffa T, Aragno M. Thermophilic bacterial communities in hot composts as revealed by most probable number counts and molecular (16S rDNA) methods. *FEMS Microbiol Ecol* 1999; 28: 141-9.
- [18] Fracchia L, Dohrmann AB, Martinotti MG, Tebbe CC. Bacterial diversity in a finished compost and vermicompost: Differences revealed by cultivation-independent analyses of PCR-amplified 16S rRNA genes. *Appl Microbiol Biotechnol* 2006; 5: 1-11.
- [19] Toyota K, Kuninaga S. Comparison of soil microbiology community between soils amended with or without farmyard manure. *Appl Soil Econ* 2006; 33: 39-48.

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